

3 **Predicting risks of invasion of macroalgae in the genus**
4 ***Caulerpa* in Florida**

5 **Christian G. Glardon · Linda J. Walters · Pedro F. Quintana-Ascencio ·**
6 **Lisa A. McCauley · Wytze T. Stam · Jeanine L. Olsen**

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9 **Abstract** There is worldwide concern about the
10 aquarium strain of the green alga *Caulerpa taxifolia*
11 (Vahl) C. Agardh that was introduced to the Medi-
12 terranean Sea in 1984. Since that time, it has
13 flourished and now covers thousands of hectares of
14 near-shore waters. More recently, aquarium strains of
15 *C. taxifolia* invaded southern California and Austr-
16 alian waters. Our goal was to evaluate potential
17 invasion of *C. taxifolia* to Florida's coastal waters.
18 We looked for evidence of *C. taxifolia*—aquarium
19 strain, as well as the present distribution of all species
20 of *Caulerpa*, in Florida's near-shore waters. We
21 surveyed 24 areas in six zones along the Floridian
22 coastline, and evaluated the association of potential
23 indicators for the presence of *Caulerpa*. Latitude,
24 presence of seagrass beds, human population density,
25 and proximity to marinas were the four variables
26 simultaneously considered. *Caulerpa taxifolia*—
27 aquarium strain was not found at any of our survey
28 locations. However, 14 species of *Caulerpa* were
29 found at 31 of the 132 sites visited. Percent correct

for our model was 61.5% for presence and 98.1% for
absence. There was a positive correlation between
Caulerpa spp. and seagrass beds and proximity to
marinas. There was a negative correlation with
latitude and human population density. The param-
eters in the logistic regression model assessing the
association of *Caulerpa* occurrence with the mea-
sured variables were then used to predict current and
future probabilities of *Caulerpa* spp. presence
throughout the state. This prediction model will
allow resource managers to focus their efforts in
future surveys.

Keywords Algae · *Caulerpa taxifolia* ·
Chlorophyta · Coastal · Invasive species ·
Prediction models

Introduction

The introduction of non-indigenous species has been
recognized as a major environmental problem for
over 100 years (e.g., Bax et al. 2001; Loope and
Howarth 2003; Barnard and Waage 2004; Perrings
et al. 2005). In the marine environment, macroalgae
in the genus *Caulerpa* are of particular concern
because of their recent expansions, ability to propa-
gate from asexual fragments, and negative impacts on
the invaded communities. *Caulerpa taxifolia*—aquarium
strain (Vahl) C. Agardh, known worldwide as the
“killer alga,” was first observed in Mediterranean

A1 C. G. Glardon · L. J. Walters · P. F. Quintana-
A2 Ascencio (✉) · L. A. McCauley
A3 Department of Biology, University of Central Florida,
A4 4000 Central Florida Blvd., Orlando, FL 32816, USA
A5 e-mail: pquintan@mail.ucf.edu

A6 W. T. Stam · J. L. Olsen
A7 Department of Marine Benthic Ecology and Evolution,
A8 Biological Centre, University of Groningen, Kercklaan 30,
A9 P.O. Box 14, 9750 AA Haren, The Netherlands

57 waters adjacent to the Monaco Oceanographic
 58 Museum in 1984 (Meinesz and Hesse 1991). It
 59 spread from an initial patch of $\sim 1^2$ m to cover
 60 hundreds of kilometers of Mediterranean coastline,
 61 where it has overgrown all native flora and fauna,
 62 impacting fisheries and tourism in coastal communi-
 63 ties (e.g., Meinesz and Hesse 1991; Relini et al. 2000;
 64 Meinesz et al. 2001; Meinesz 2002). In 2000,
 65 *C. taxifolia*—aquarium strain was discovered in two
 66 lagoons in southern California (Jousson et al. 2000)
 67 and in New South Wales, Australia in the Port
 68 Hacking, Careel Bay and Lake Conjola regions (Grey
 69 2001; Wiedenmann et al. 2001; Millar and Talbot
 70 2002; Schaffelke et al. 2002). Eradication was suc-
 71 cessful in California (R. Woodfield, personal
 72 communication), while its spread continues in
 73 Australia (Millar 2004). It is suspected that human
 74 activities, either boating or aquarium releases, were
 75 responsible for all invasions (e.g., Meinesz 1999;
 76 Raloff 2000; Millar and Talbot 2002).

77 Several other *Caulerpa* species may be able to
 78 outcompete native macrophytes and create monospe-
 79 cific beds (Verlaque and Fritayre 1994; Piazzini et al.
 80 2001; Piazzini and Ceccherelli 2002, 2006). Since
 81 1990, *C. racemosa* var. *cylindracea* has been rapidly
 82 spreading and dramatically expanding throughout the
 83 Mediterranean Sea and Canary Islands (Verlaque
 84 et al. 2000, 2003; Ruitton et al. 2005). Similarly, in
 85 2001, non-native *C. brachypus* created concern along
 86 the east coast of south Florida, where it was locally
 87 abundant, displacing native flora and fauna (Schrope
 88 2003; Jacoby et al. 2004; SFER 2005). *Caulerpa*
 89 *brachypus* spread north into the Indian River Lagoon
 90 system, which was consistent with prevailing coastal
 91 currents, but it has not been reported in west Florida
 92 (Schrope 2003). Most *C. brachypus* did not survive
 93 Hurricanes Frances and Jeanne that battered Florida
 94 in the late summer of 2004 (B. LaPointe, personal
 95 communication).

96 Florida's coastline closely matches environmental
 97 conditions of other areas invaded by *C. taxifolia*—
 98 aquarium strain and the risk is significant that this
 99 state will be invaded in the near future. Non-invasive
 100 *C. taxifolia* has a lower lethal temperature of 14°C,
 101 while mortality of the aquarium strain of *C. taxifolia*
 102 from Mediterranean waters is 7°C (Komatsu et al.
 103 1994; Ramey 2001). Seagrasses are frequently asso-
 104 ciated with various *Caulerpa* species. In some cases,
 105 the presence of one can facilitate the other through

106 stabilization of the substrate (Williams 1984, 1990;
 107 Smith and Walters 1999; Magalhaes et al. 2003). In
 108 disturbed areas, however, the situation is different
 109 (e.g., Stafford and Bell 2006). In a number of areas
 110 along the Mediterranean coastline, *C. taxifolia* was
 111 able to outcompete *Posidonia oceanica* (Chisholm
 112 and Jaubert 1997; Villele and Verlaque 1994) and
 113 *Cymodocea nodosa* (Relini et al. 1998a, b, c).

114 High human population density may increase or
 115 decrease the probability of a marine macrophyte
 116 invasion. Boaters increase the potential and fre-
 117 quency of transport via fragments and propagules in
 118 ballast tanks, live wells, or attached to propellers and
 119 hulls. Along the French Mediterranean coast, all areas
 120 colonized by invasive *C. racemosa* var. *cylindracea*
 121 (Sonder) Verlaque, Huisman et Boudouresque were
 122 associated with human activities and over 40% were
 123 in fishing areas (Ruitton et al. 2005). Releases of
 124 aquarium organisms into storm drains or local
 125 waterways by well-meaning hobbyists will also
 126 increase as the population density and number of
 127 aquaria increases. Although the aquarium strain of
 128 *C. taxifolia* is banned from importation and interstate
 129 transport in the USA, other species of *Caulerpa*
 130 remain very popular with hobbyists. For example,
 131 non-invasive strains of *C. taxifolia* and 12 additional
 132 species of *Caulerpa* are readily available via local
 133 and Internet retailers as well as Internet auction sites
 134 (Walters et al. 2006; Zaleski and Murray 2006).
 135 Coastal population pressure also holds a higher
 136 potential for greater pollutant loads, freshwater and
 137 nutrient run-offs; these may prevent or increase
 138 algal growth (Morand and Merceron 2005). In the
 139 Mediterranean, *C. taxifolia*—aquarium strain was
 140 concentrated in zones with extensive development
 141 (Madl and Yip 2005).

142 Considering the length of Florida's shoreline and
 143 the economic and environmental importance of these
 144 waters, it is urgent to be prepared for a human-
 145 mediated introduction of *Caulerpa*. Our goal was to
 146 determine locations that are most susceptible to
 147 *Caulerpa* invasion by aquarium releases or boating
 148 activities, and that would be most suitable for
 149 recruitment of species of the genus *Caulerpa*. Being
 150 able to concentrate on areas that are more at risk
 151 would greatly help prevention and eradication efforts.
 152 Two questions are fundamental to this goal: (1) What
 153 habitat(s) are most suitable for *Caulerpa*?, (2) What
 154 areas are most likely to be invaded, especially if

155 home aquarium releases or recreational boating are
156 involved?

157 Methods

158 We used a stratified sampling design to assess the
159 current distribution of *Caulerpa* spp. along the
160 Florida shoreline and then to test the association of
161 *Caulerpa* spp. occurrence with variables allowing us
162 to evaluate its risks of invasion (see below). We
163 chose to stratify the Florida shoreline to reflect the
164 latitudinal and longitudinal variation in water tem-
165 perature, seagrass presence/absence, local human
166 population density, and the presence/absence of boat
167 marinas. All GIS data were downloaded from the
168 Florida Geographic Data Library (2006).

169 We obtained bi-monthly sea surface temperature
170 for the Floridian coastline (2001–2004). Data were
171 available as a grid of 14 km per side (Comprehensive
172 Large Array-Data Stewardship System: www.class-
173 noaa.gov). We transferred the temperature data along
174 the coastline to an Excel spreadsheet (207 pixels) and
175 performed a non-metric, multi-dimensional scaling
176 ordination for each summary temperature (monthly,
177 seasonal, and annual) using PC-Ord (McCune and
178 Grace 2002). The mean temperature for January each
179 year had the largest range of temperatures and sorted
180 into six distinct groups (Fig. 1). We used gaps or
181 switches in the values of the final single ordination
182 axes to define the groups. From the western extreme
183 of the Floridian coastline, the first zone ended at
184 85°W longitude; the second zone ended near Tampa
185 at 28°N; the third zone went down to Key West, FL

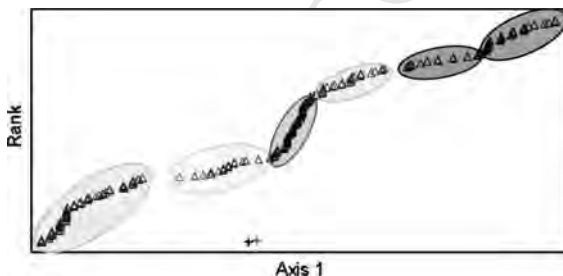


Fig. 1 Single axes of a non-metric, multi-dimensional scaling ordination for temperature data around the coastline of Florida (MPC-Ord MjM Software Design). The six groups show the six different zones of different temperature range used as the basis for the stratification of the state

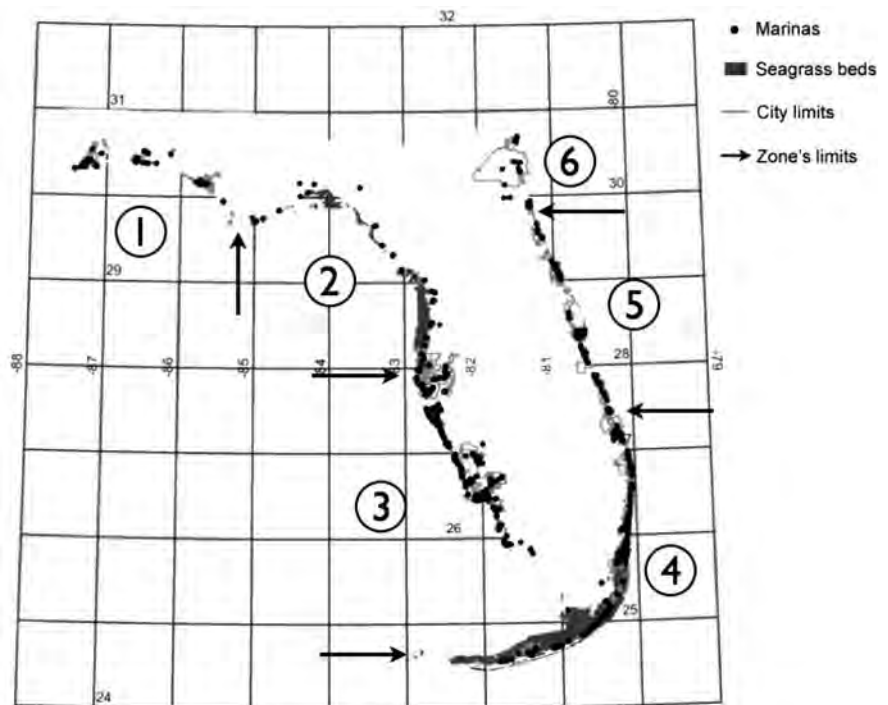
186 Keys; the fourth went from Key West up the east
187 coast to longitude 27°N; the fifth up to the longitude
188 29°N; and the sixth up to the Georgia border (Fig. 2).
189 In our analysis we used latitude instead of temper-
190 ature because there was a significant correlation
191 between temperature and latitude (-0.94 , $n = 207$);
192 this allowed us to directly map the model results.

193 We considered all seagrasses occurring in Florida,
194 including *Halodule wrightii*, *Syringodium filiforme*,
195 *Thalassia testudinum*, *Halophila johnsonii*, *Halophila*
196 *decipiens*, *Halophila engelmannii*, and *Ruppia mari-*
197 *time* (Virnstein and Morris 1996). GIS coverages of
198 their distribution around the coastline of Florida were
199 available from a number of sources: Florida Fish and
200 Wildlife Conservation Commission, Florida Marine
201 Research Institute and Coastal and Marine Resource
202 Assessment (Fig. 2). We arbitrarily chose 25,000
203 inhabitants within the city limits to be the cut-off
204 between low human-impacted and high human-
205 impacted areas. Boat traffic transporting species from
206 one area to another makes marinas prone to becoming
207 primary invasion sites (Boudouresque et al. 1995;
208 Loope 2004). Docks are also frequently areas where
209 people have easy access to marine environment,
210 making them logical locations for disposing of
211 unwanted aquarium plants and animals. Marinas are
212 well represented around the state of Florida (FGDL
213 2006) (Fig. 2).

214 Bathymetry was not chosen as a variable because
215 our sampling was restricted to depths of <10 m. This
216 should, however, not pose a problem since most
217 native *Caulerpa* spp. occur above 20 m (Littler et al.
218 1989). Thibaut et al. (2004) also reported higher
219 biomasses of *C. taxifolia*—aquarium strain between 6
220 and 10 m. However, invasive *C. racemosa* var.
221 *cylindracea* along the French Mediterranean coast
222 was found primarily between 10 and 35 m (Ruitton
223 et al. 2005). Substrate type (e.g., grain size) and
224 shoreline vegetation were not chosen because of the
225 lack of support from the literature that would give an
226 eventual correlation with marine species occurrence.
227 Water chemistry was not used because data was
228 limited to certain stations and did not cover the entire
229 coastline.

230 Using ArcMap 9.1, we combined data layers of
231 five variables: latitude and longitude (as continuous
232 variables), and seagrass presence/absence, local low/
233 high human population density, and presence/absence
234 of boat marinas (as categorical variables). A line data

Fig. 2 Stratification of the state of Florida into six zones of different temperature ranges (1 = 16–31.5°C, 2 = 12.5–31.5°C, 3 = 17–31.5°C, 4 = 23–31 C, 5 = 20–30°C, and 6 = 12.5–31.5°C), and locations of seagrass beds, marinas, and areas with >25,000 people per square mile



235 layer of the Florida coastline was buffered by 3 km in
 236 order to integrate lagoons and estuaries. We concen-
 237 trated only on recruitment that might have resulted
 238 from a release from a marina. Hence, the data layer
 239 for marinas had a buffer of 2 km around each marina.
 240 The spreading of *Caulerpa* fragments showed that
 241 there is a gradient of natural fragment dispersal over
 242 short distances (several hundred meters, Hill et al.
 243 1998). City/town (with city limits) and seagrass
 244 presence/absence data layers were merged with the
 245 buffered marinas data layer. The marina/seagrass/city
 246 data layer was merged with the buffered temperature
 247 zones data layer and the merged data layer was
 248 clipped to the extent of the zones layer (3 km around
 249 the entire coastline).

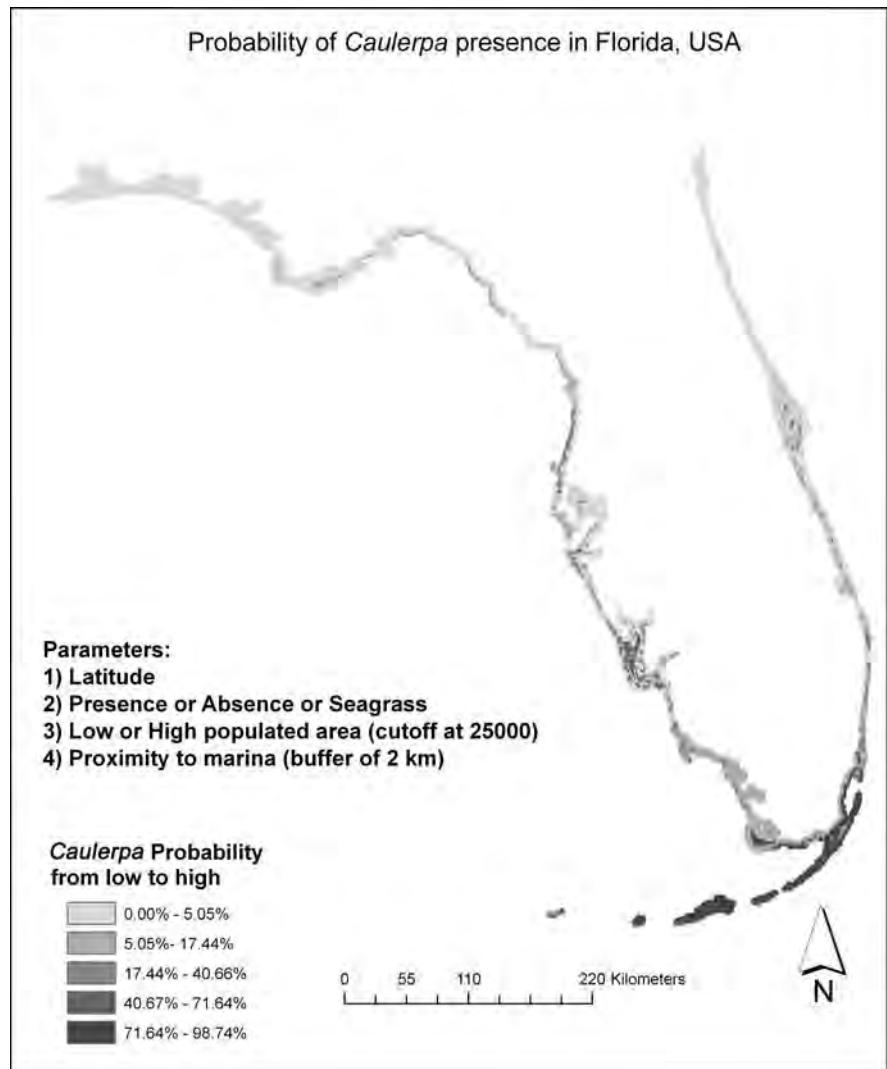
250 We used this map to delimit the areas that
 251 corresponded to all the possible combinations of
 252 parameters and for choosing our survey locations
 253 (Fig. 2). Within each delimited area for the 48
 254 different associations of variables (six zones \times three
 255 replicates $\times 2^3 = 144$ locations), we randomly
 256 choose three sample points after eliminating areas
 257 that were inaccessible (i.e., US Air Force Bases).
 258 Exact GPS coordinates were used to access each
 259 location using a handheld Garmin e-Trex GPS
 260 receiver (accuracy < 14 m). Once on site, we

261 snorkeled over a rectangular area that extended 261
 262 20 m perpendicular from the shoreline and 100 m 262
 263 parallel the shoreline (centered on the GPS point), 263
 264 and recorded the presence of each species of 264
 265 *Caulerpa*. 265

266 We located and surveyed 132 points of the 144 266
 267 points anticipated, and entered the data in an Excel 267
 268 spreadsheet. Twelve points were not considered 268
 269 because the sixth temperature zone (northeast Florida) 269
 270 did not have any seagrass. During the fieldwork, we 270
 271 confirmed the association of the point with the 271
 272 anticipated state of the variables in each location. 272

273 We used multiple logistic regression models (all 273
 274 possible nested models; SPSS 11.0, MacOSX) to test 274
 275 the association of *Caulerpa* with the four independent 275
 276 variables: GPS latitude coordinate, presence of sea- 276
 277 grass (present/absent), population density (high/low), 277
 278 and proximity to marina (with the 2 km buffer zone/ 278
 279 outside the 2 km buffer zone). We next used the 279
 280 Akaike's Information Criterion (AIC) to select the 280
 281 "best" multiple linear regression model (Burnham and 281
 282 Anderson 2002). The parameters of this model were 282
 283 then used to predict the probability of *Caulerpa* 283
 284 occurrence across the Florida shoreline based on a 284
 285 multi-layered grid. We created a 1,000 \times 1,000 m² 285
 286 cell grid from the initial data layer described above 286

Fig. 3 Probability of *Caulerpa* spp. presence along the coastline of Florida based on logistic regression using latitude, seagrass presence/absence, population density, and marina proximity. The best model was selected using Akaike's Information Criterion



287 and the centroid of each cell was used to assign the
 288 environmental variables for that cell. We predicted
 289 the probability of *Caulerpa* presence in each cell using
 290 the parameters of the best logistic regression model and
 291 created a graduated color map for the entire state
 292 showing these probabilities for the entire coast of
 293 Florida (Fig. 3). The probability ranges were chosen to
 294 reflect the maximum heterogeneity of the data.

295 Results

296 We found *Caulerpa* in 31 of 132 surveyed zones sites,
 297 including *C. prolifera* (15 occurrences), *C. sertulario-*
 298 *ides* (10), *C. paspaloides* (9), *C. mexicana* (9),

C. cupressoides (7), *C. ashmeadii* (5), *C. lanuginosa* 299
 (5), *C. verticillata* (3), *C. racemosa* (3), and 300
C. microphysa (1) (Fig. 4). No *C. taxifolia* was 301
 observed in our surveys. Among the 31 sites where 302
Caulerpa species were found, 24 were in seagrass 303
 beds. Eighteen of the 31 sites where *Caulerpa* species 304
 were found were in locations with low populations 305
 (<25,000 inhabitants). Eighteen of the 31 sites with 306
Caulerpa spp. were within 2 km of a marina 307
 (<2 km). Local species richness of *Caulerpa* 308
 increased as latitude decreased (an inverse relation- 309
 ship), and with the presence of seagrass, and 310
 decreased with human density ($r^2 = 0.378$, $n = 132$, 311
 $P < 0.0001$, and $y = -11.733 + 346.255/\text{GPS} +$ 312
 $0.647 \times \text{seagrass} - 0.604 \times \text{population density}$). 313

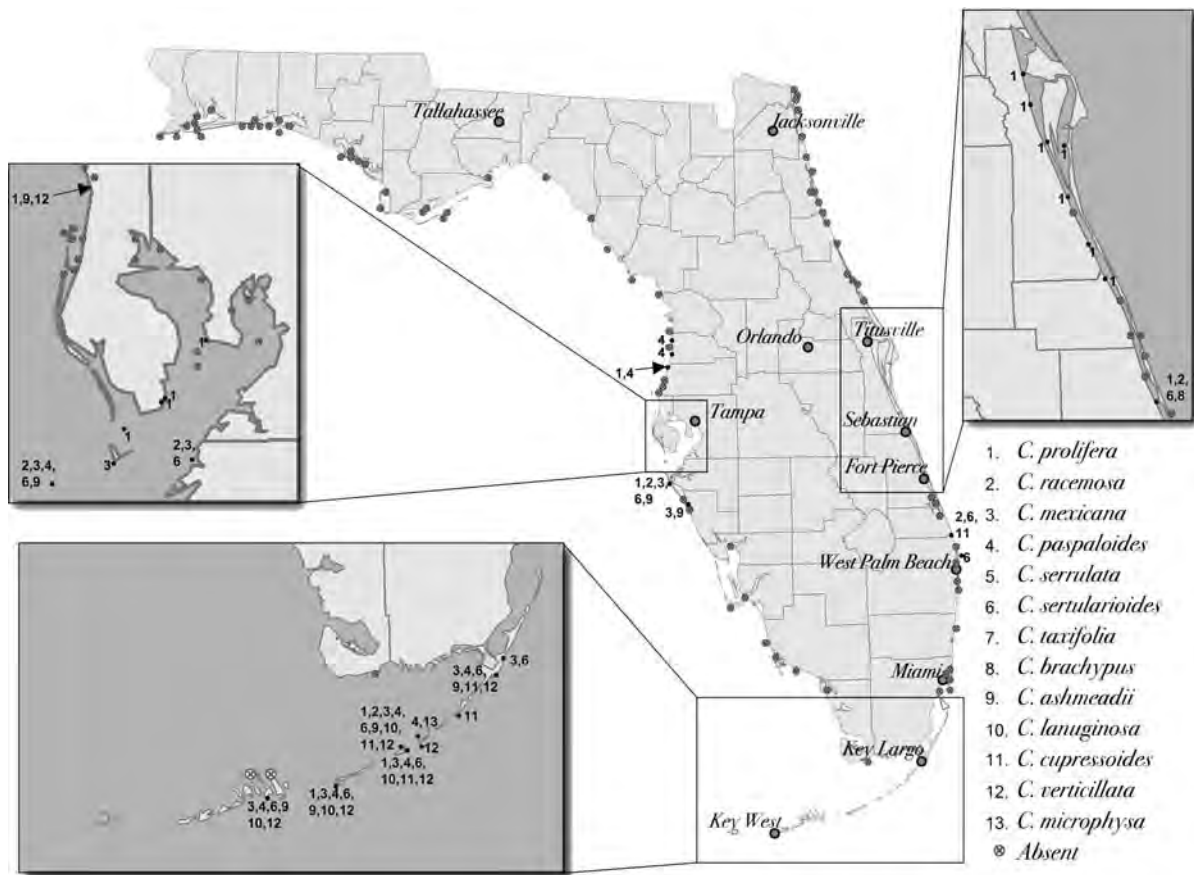


Fig. 4 Current distribution of *Caulerpa* by species around the coast of Florida. Numbers at each location list all species of *Caulerpa* found at that site and match the list on the lower right of the figure

314 Combined, the probability of *Caulerpa* occurrence
 315 increased as the latitude decreased, with the presence
 316 of seagrass beds, in sites with low density of human
 317 populations, and in close proximity to marinas
 318 (Table 1). An assessment of the possible logistic
 319 regression models with these variables using
 320 Akaike's information criteria indicated that the full
 321 model including all four parameters was the best
 322 (weight of 0.83; Table 1). Percent correct for *Caulerpa*
 323 presence and absence for this model were 61.5
 324 and 98.1%, respectively (Table 2). We used the
 325 parameters of this model to predict the occurrence of
 326 *Caulerpa* along Florida shoreline (Fig. 3).

327 We also assessed the logistic regression models
 328 for single species. Separately, *Caulerpa prolifera*,
 329 *C. paspaloides*, *C. mexicana*, and *C. sertularioides*
 330 had significant correlations ($P < 0.05$) with latitude
 331 (*C. prolifera*: Table 3). Percent correct presence and
 332 absence for the model with latitude were 13.3

and 97.4%, respectively, for *C. prolifera*; 55.6 and 333
 99.2%, respectively, for *C. paspaloides*; 77.8 334
 and 98.4%, respectively, for *C. mexicana*; and 50 335
 and 99.2%, respectively, for *C. sertularioides*. 336

Discussion 337

338 Our model provides information about current loca- 338
 339 tions of *Caulerpa* species and of potential suitable 339
 340 zones for recruitment, making it an important 340
 341 conservation tool. In agreement with prior informa- 341
 342 tion, our model indicates that *Caulerpa* occurs 342
 343 preferentially in warmer waters and in habitats with 343
 344 seagrass. We differed from findings by Madl and Yip 344
 345 (2005) and Ruitton et al. (2005) that *Caulerpa* was 345
 346 associated with extensive human activities, as we 346
 347 found *Caulerpa* most frequently in areas of low 347
 348 human density areas. These differences may be 348

Table 1 Summary of Akaike’s information criteria and associated statistics for the nested logistic regression models of *Caulerpa* occurrence data in the Florida peninsula

Model	−2 Log likelihood	Par	AIC	AIC dif	Weights
GPS, seagrass, marina, population	64.11	6	76.1	0	0.83
GPS, seagrass, marina	69.49	5	79.5	3.4	0.15
GPS, seagrass, population	74.53	5	84.5	8.4	0.01
GPS, seagrass	79.74	4	87.7	11.6	0.002
Marina, seagrass, population	85.91	5	95.9	19.8	<0.001
Marina, seagrass	92.78	4	100.8	24.7	<0.001
Seagrass, population	92.78	4	100.8	24.7	<0.001
GPS, marina, population	92.85	5	102.8	26.7	<0.001
GPS, marina	96.20	4	104.2	28.2	<0.001
Seagrass	99.04	3	105.0	28.9	<0.001
GPS, population	100.81	4	108.8	32.7	<0.001
GPS	104.26	3	110.3	34.1	<0.001
Marina, population	121.02	4	129.0	52.9	<0.001
Marina	126.10	3	132.1	56.0	<0.001
Population	126.10	3	132.1	56.0	<0.001

Table 2 Summary statistics for the best models for all species of *Caulerpa* and single species occurrences

Model	Model parameters					
	Percent correct	Constant	GPS	Seagrass	Popdens	Marina
GPS, seagrass, marina, population (all species)	90.9	21.029	−0.918	3.741	−1.491	2.127
GPS, seagrass, marina (<i>prolifera</i>)	87.9	7.603	−0.427	2.081		1.428
GPS (<i>paspaloides</i>)	94.7	24.85	−1.008			
GPS (<i>mexicana</i>)	97	41.026	−1.628			
GPS (<i>sertularioides</i>)	95.5	32.203	−1.282			

Table 3 Summary of Akaike’s information criteria and associated statistics for the nested logistic regression models of *Caulerpa prolifera* occurrence data in the Florida peninsula

Model	−2 Log likelihood	Par	AIC	AIC dif	Weights
GPS, seagrass, marina	71.15	5	81.1	0	0.682
GPS, seagrass	76.21	4	84.2	3.065	0.147
Marina, seagrass	76.82	4	84.8	3.671	0.109
GPS, marina	80.18	4	88.2	7.031	0.020
Seagrass	81.00	3	87.0	5.849	0.037
GPS	85.01	3	91.0	9.86	0.005
Marina	89.65	3	95.6	14.505	0.0005

349 species-specific as we considered only native species
 350 and they focused on highly invasive strains. Unsur-
 351 prisingly, close proximity to marinas was positively
 352 correlated with *Caulerpa* presence.

Our analysis showed that the presence of seagrass
 was the best predictor of the presence of *Caulerpa*
 among the four variables. About 24 of 31 sites with
Caulerpa had seagrass, regardless of the association

353
 354
 355
 356

with all other parameters. Seagrasses depend on sediment-based decomposition of organic matter and elemental recycling and are prone to human disturbances (McRoy and McMillan 1977; Klug 1980; McRoy and Lloyd 1981; Thayer et al. 1975; Lewis 1987; Livingston 1987; Williams 1990). Seagrasses obtain a large fraction of their nutrients from the sediment via roots, while leaf uptake is considered of secondary importance (Pedersen and Borum 1993; Ceccherelli and Cinelli 1997). *Caulerpa* can utilize both sediment and water column nutrients (Williams 1984), which may account for the strong correlation. *Caulerpa* is endemic in tropical and subtropical regions around the world and latitude was a significant predictor of its native occurrence (Creese et al. 2004; Zaleski and Murray 2006; Stam et al. 2006). However, Silva (2002) mentioned that this genus can also grow in locations as high as 34°N. Although Florida lies between latitude north 24 and 30°N and thus, has the potential for *Caulerpa* recruitment along its entire coastline, we found that *Caulerpa* species richness and occurrence was negatively correlated to latitude. Other physical factors may explain this pattern. The large tidal regime, large expanses of bare sand and wave energy on the northern Atlantic seaboard of Florida and the Panhandle region of Florida (Gulf of Mexico) that prevent seagrasses from establishing may also prevent *Caulerpa* spp. recruitment (L. Morris personal communication). Unstable substrates such as ripple-marked sediments and shallow rocky shores exposed to strong wave action are some of the rare locations where *C. taxifolia*—aquarium strain can not become established, while protected areas, such as lagoons or coral reefs, offer better potential for recruitment (de Vaugelas et al. 1999). In the Panhandle, many other survey locations were in or close to estuaries and bays, such as Pensacola Bay and Choctawhatchee Bay near Fort Walton Beach, West and East Bays near Panama City and Apalachicola Bay near Apalachicola. These sites were characterized by high fresh water runoff, as well as higher population densities. The low salinity in these areas, often <10 ppt, is lethal for *Caulerpa* (Madl and Yip 2005), and could also account for the lack of *Caulerpa* at these sites. South of 29°N, *Caulerpa* was found in protected environments such as lagoons, seagrass beds, or attached to highly structured surfaces, such as jetties or hard corals.

Proximity to marinas was the next most important variable correlated to *Caulerpa* spp. occurrence. The

incidence of native *Caulerpa* around marinas suggests a higher risk of recruitment of native or non-native *Caulerpa* if disposed of at marina locations. Because of the favorable habitat and its easy access to humans, these coastal waterways can be areas where species that are the object of trade for home aquarium industry have a significant probability of successful release in the wild (Loope 2004; Padilla and Williams 2004; Walters et al. 2006; Stam et al. 2006). Marinas are also areas where boat traffic favors the spread of species through ballast water, live wells for bait, or through fragments attached to hulls, anchors or traps (Loope 2004; Madl and Yip 2005). Approximately 1 year after our surveys were completed (August 2006), we received inquiry from a scientist working in Destin Harbor, FL (Panhandle region) (J. Fry, personal communication). Their group had discovered two dense beds of *C. sertularioides* in ~3 m of water near the local marina. They had not previously recorded this species in this location. We had searched nearby waters (<1 km away) in August 2005 and found no evidence of *Caulerpa*. So, we now have our first evidence to suggest that marinas are good locations for *Caulerpa* to enter Florida waters. *Caulerpa sertularioides* is native to Florida, so eradication is unlikely unless it proves to be a new strain.

Human population density was negatively correlated to overall *Caulerpa* presence. Heavily populated areas (>25,000) might be areas with too many disturbances for *Caulerpa* spp. recruitment. During our surveys, we often observed these areas to have anoxic substrates and have high turbidity. These conditions do not favor recruitment or survival of either angiosperms or macroalgae (Plus et al. 2003). However, Chisholm et al. (1997) showed that *C. taxifolia*—aquarium strain proliferated in areas of urban wastewater pollution. This might be a unique feature of the invasive, aquarium strain of *C. taxifolia*.

Caulerpa cupressoides, *C. ashmeadii*, *C. lanuginosa*, *C. verticillata*, and *C. microphysa* were only observed in the Florida Keys (Fig. 4). None of these species were significantly correlated with any of the tested variables. Small sample size is likely to be the major reason why no inference could be made (Hirzel and Guisan 2002). *Caulerpa prolifera*, *C. mexicana*, *C. paspaloides*, *C. racemosa*, and *C. sertularioides* were more likely to settle further north than other species. These species showed a negative correlation with latitude.

Caulerpa taxifolia—aquarium strain has a lower lethal temperature limit than the native strain, 7 and 14°C, respectively (Komatsu et al. 1994; Ramey 2001). Thus, the potential distribution of the aquarium strain based on temperature extends throughout the entire Florida coastline and should extend further north along the Atlantic seaboard than any native species of *Caulerpa*. Although absent in our zone 6, *Caulerpa* species are present further north and can grow in locations like the Onslow Bay, North Carolina, at latitude 34°N (Silva 2003). This suggests that areas north of North Carolina that are too cold for native *Caulerpa* may be suitable for establishment of *C. taxifolia*—aquarium strain and resource managers should be aware of this.

Our data indicate that latitude, presence of seagrass, human population density, and proximity to marinas successfully predict the occurrence of *Caulerpa* species along the Florida coastline and can be a useful tool to select zones for survey that would be more likely to be invaded by *Caulerpa*. It now needs to be combined with effective monitoring programs that can lead to rapid identification and eradication. Otherwise, the number of invasions and their subsequent effects will only increase (Bax et al. 2001). Also, we must consider that climate change is likely to shift the distribution of suitable areas for many species, including *Caulerpa* (Williams and Schroeder 2004). Thus, this model, as any other model, is a temporary tool in need of constant adaptation to new environmental and human factors.

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